

Mobilization of Heavy Metals in Contaminated Soils induced by Bioaugmentation of *Shewanella xiamenensis* HM14

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(Received: August 11 2014, Revised: August 13 2014, Accepted: August 24 2014)

A bacterial strain with the potential ability to solubilize heavy metals was isolated from heavy metal contaminated soils collected from abandoned mines of Boryeong area in South Korea. The bacterial strain with the highest degree of metal resistance was shown to have close proximity with *Shewanella xiamenensis* FJ589031, according to 16S rRNA sequence analysis, and selected for investigating the mobilization of metals in soil or plant by the strain. The strain was found to be capable of solubilizing metals both in the absence and in the presence of metals (Co, Pb and Cd). Metal mobilization potential of the strain was assessed in a batch experiment and the results showed that inoculation could increase the concentrations of water soluble Co, Pb and Cd by 48, 34 and 20% respectively, compared with those of non-inoculated soils. Bacterial-assisted growth promotion and metal uptake in sunflower (*Helianthus annuus*) was evaluated in a pot experiment. In comparison with non-inoculated seedlings, the inoculation led to increase the growth of *H. annuus* by 24, 18 and 16% respectively in Co, Pb and Cd contaminated soils. Moreover, enhanced accumulation of Co, Pb and Cd in the shoot and root systems was observed in inoculated plants, where metal translocation from root to the above-ground tissues was also found to be enhanced by the strain. Plant growth promotion and metal mobilizing potential of the strain suggest that the strain could effectively be employed in enhancing phytoextraction of Co, Pb and Cd from contaminated soils.

Key words: Mobilization, *Shewanella xiamenensis* HM14, Bioaugmentation, Sunflower

Effect of inoculation with *Shewanella xiamenensis* HM14 on accumulation and translocation of Co, Pb and Cd in *Helianthus annuus*.

Metal	Treatment	Metal content (mg/kg dry weight)		Bioconcentration Factor (BCF) ^a	Translocation Factor (TF) ^b
		Shoot	Root		
Co	control	19.24 (± 3.04)	81.23 (± 5.84)	0.456	0.237
	with strain	26.57 (± 3.57)	109.71 (± 6.27)	0.548	0.242
Pb	control	11.82 (± 0.87)*	68.53 (± 9.36)	0.343	0.172
	with strain	14.13 (± 0.91)*	85.26 (± 7.37)	0.426	0.166
Cd	control	12.69 (± 6.43)	48.57 (± 6.25)*	0.243	0.261
	with strain	18.42 (± 12.25)	56.96 (± 7.24)*	0.285	0.323

^aBCF=metal concentration ratio of plant roots to soil; ^bTF=metal concentration ratio of plant shoots to roots.

Values are means ($n=3$) ± standard deviation. Within each column, the means indexed by * are not significantly different at $p > 0.05$ between inoculated and non-inoculated plants according to Duncan's multiple range test.

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§Acknowledgement: This study was supported by a grant from the research project of National Institute of Horticultural & Herbal Science, Rural Development Administration, Republic of Korea.

Introduction

Metal contamination of soils has become one of the most significant environmental problems today. Excessive metal uptake by crop plants from the contaminated agricultural lands can result in decreased crop yield due to the inhibition of plant metabolic processes (Singh and Aggarwal, 2006). Apart from the metals with unknown biological functions (Cd, Cr, Pb, Co, Ag, Se, and Hg), essential elements (Fe, Mn, Zn, Cu, Mg, Mo, and Ni) also keep accumulating in agricultural soils by means of wastewater irrigation, animal manures and sewage sludge application, use of fertilizer and agrochemicals (Thomas *et al.*, 2012). In the toxicological point of view, the essential elements are also important, because, at higher concentrations they too can be toxic to plants as well as to dietary intake levels (Karavoltzos *et al.*, 2002).

With the continuous addition of undesirable metals into the environment, remediation of contaminated soils receives increasing attention (Yeh and Pan, 2012). However, due to the fact that metals are not easily mineralized, remediation of the contaminated soils is always considered being a demanding exercise (Rajkumar *et al.*, 2008). Depending on the resource availability, severity of the problem, nature of the metals and contaminated soil, different methods have been employed in restoring the contaminated lands (Arunakumara *et al.*, 2013). In this context, systematic technologies such as bioremediation, physical/chemical remediation and integrated remediation are among the widely used techniques (Luo, 2009). However, the physical and chemical methods such as physical separation, acid leaching or electrochemical processes, are considered to be ineffective because of high cost, low efficiency, and destruction of soil structure and fertility (Jing *et al.*, 2007). In contrast, phytoremediation, a method which uses plants to extract, sequester and detoxify pollutants has received considerable attention (Arunakumara, 2011). However, the wider application of the technology has been restricted due to the limitations such as low soil thickness that can be treated, low translocation rate of metals from roots to shoots, and the slowness of the treatment (Lebeau *et al.*, 2008).

The amount of heavy metals uptake in plants varies with the mobility and the concentration of metals in soil (Chen *et al.*, 2010) and the interface between soil microbes and plant roots (rhizosphere) is displayed to have a great influence on the uptake of nutrients as well as on the decrease of metal toxicity (Ryan *et al.*, 2009). Since soil microbes could alter the metal status of the soil (Fazal and Bano, 2010), exploitation of such microbes to reduce the metal toxicity to plants is worth investigating (Rajkumar and Freitas, 2008). In this context, some metal resistant bacterial strains were proved exceptional at enhancing the growth of the host plant through different mechanisms such as the production of plant growth promoting substances, nitrogen fixation and phosphate solubilization

(Hemambika *et al.*, 2013). As reported by Rajkumar *et al.* (2008), heavy metal tolerance of the microbes may be attributed to one or several mechanisms including exclusion, active removal, biosorption, and precipitation or bioaccumulation of metals both in external and intracellular spaces. Therefore, microbes having remarkable metal tolerance and plant growth-promoting abilities could play a significant role in remediation of metal-contaminated soils, because bioaugmentation with such microbes could promote phytoextraction of metals (Prapagdee *et al.*, 2013). In the present study, we isolated heavy metal resistance bacterial strains from metal-contaminated soils and the strain with the highest degree of metal resistance was employed in (i) assessing the potential of mobilization of Co, Pb and Cd, and (ii) evaluating the effects of inoculation with the selected strain on plant growth and uptake of Co, Pb and Cd by *Helianthus annuus* (sunflower).

Materials and Methods

Isolation of heavy metal alleviating bacteria Heavy metal contaminated soils were collected from the sediment tailing in abandoned mines of Boryeong area, South Korea. Aliquots of serially diluted soil samples were spread on Tryptic soy broth (TSB, 30 g L⁻¹) agar media which was adjusted to pH 6.5 ± 0.1. The colonies on agar plate were purely isolated by repeated sub culturing at 30°C for 5 days. The preculture of each colonies was inoculated in 50 ml TSB containing 0.2 g -0.4 g L⁻¹ of CoCl₂·6H₂O, 2PbCO₃·Pb(OH)₂, and Cd(NO₃)₂ at 30°C for 7 days on a horizontal shaker at 150 rpm and the culture filtrate was recovered after centrifugation at 8000×g to isolate a metal alleviating bacteria. To determine aqueous elemental concentrations including Co, Pb, and Cd, the culture filtrates were filtered through a 0.45 μm nylon syringe filter (Watman, England) and acidified with HNO₃ to minimize an interference by organic matters. The metal content of samples were analyzed using inductively-coupled plasma optical emission spectroscopy (ICP-AES, Optima 5300DA, PerkinElmer, USA) to estimate the metal concentration alleviated by each bacterium in medium and to select the bacterial strain showing the highest degree of metal alleviation.

Strain identification The partial sequencing of 16S rRNA for the bacterial strain was done with the help of DNA sequencing service, SOLGENT, Daejeon, South Korea using universal primers, 27F (5'-AGAGTTTGATCCTGGCTCAG-3') and 1492R (5'-GGTTACCTTGTTACGACTT-3'). The online program BLAST was used in identifying the related sequences with known taxonomic information available at the databank of NCBI (<http://www.ncbi.nlm.nih.gov/BLAST>). A Phylogenetic tree was constructed using CLUSTAL X program (Thompson *et al.*, 1997), which included sequence alignment by neighbor joining method (Saitou and Nei, 1987) and

maximum parsimony using the MEGA4 program (Kumar *et al.*, 2001). Grouping of sequences was based on confidence values obtained by bootstrap analysis of 1,000 replicates. Gaps were edited in the BioEdit program and evolutionary distances were calculated using Kimura two parameter model. Reference sequences were retrieved from GenBank under the accession numbers indicated in the trees.

Effect of heavy metals on bacterial growth TSB medium supplemented with heavy metals (Co, Pb and Cd) at the concentration of 200 mg L⁻¹ was inoculated with bacterial suspension (10⁶ colony forming units ml⁻¹) and incubated with continuous shaking at 30°C for 5 days. Optical density of culture supernatant was measured at definite time intervals using UV spectrophotometer at 660 nm to estimate the cell growth.

Assay of heavy metal contents in medium Bacterial culture having 10⁶ colony forming units ml⁻¹ (2 days old) was inoculated in sterilized liquid TSB medium (250 ml) supplemented with different heavy metals (Co, Pb and Cd) at the concentration of 200 mg L⁻¹ and incubated with continuous shaking at 30°C. Sterilized liquid TSB medium without supplemented with heavy metals was served as a control. A sample (10 ml) of each culture and control were taken and centrifuged at 12000 ×g for 15 min. The clear supernatant was used in determining the pH and amount of the metals remaining in the medium.

Effect of bioaugmentation on growth and metal uptake by *H. annuus* A pot experiment was conducted under greenhouse conditions at the College of Agriculture, Chungnam National University in April 2014. Several locations of soils collected from abandoned mines of Boryeong-gun as contaminated soil and a waste button mushroom compost in Buyeo-gun area, Chungchungnam-do, South Korea, were respectively mixed with the ratio of 1:1, air dried and sieved (2 mm). Sterilized forest soil (by steaming at 100°C for three consecutive days) was amended with aqueous solutions of different heavy metals (Co, Pb and Cd) to achieve the final concentrations of 200 mg/kg soil. They were then kept for 2 weeks in a greenhouse for metal stabilization and used in filling the plastic pots (25 cm diameter, 35 cm height). Seeds of *Helianthus annuus* were surface sterilized by immersing in alcohol (70%) for 40 s, NaClO (1.0%) for 15 min followed by rinsing several times with sterile distilled water. Seeds sown in germination trays containing sterilized non-contaminated soil were provided with 14/10 light/dark regime and kept at 25°C for germination. Bacterial cultures grown under standard conditions for 2 days were harvested by centrifugation at 12000 ×g at 15 min. The cells were washed twice with sterile distilled water and resuspended in biological saline (0.85% KCl) to be used in inoculation. Three weeks old seedlings were carefully uprooted from the germination bed and their roots were dipped in the

bacterial culture (10⁹ colony forming units ml⁻¹) for 2 h. They were transplanted into the plastic pots (five plants/pot) containing 300 g of metal contaminated or non-contaminated soil and allowed to grow at 25°C and 14/10 light/dark regime. The average pH of soil at the time of planting was recorded as 6.65. Three weeks later, the plants were carefully uprooted and cleaned the root surface thoroughly with distilled water. As growth parameters, fresh and dry biomass was measured and accumulation of metals in plant biomass was quantified as described by Freitas *et al.* (2004). Each treatment had three replicates.

Mobility of the metals in soil The impact of bacterial inoculation on the mobility of metals in soil was investigated under laboratory conditions with 50 ml scaled polypropylene centrifuge tubes. The bacterial strain transferred into 100 ml flasks containing TSB was cultured aerobically on a rotating shaker (150 rpm) at 30°C until reaching the final concentration of 10⁸ colony forming units ml⁻¹. The bacterial cells were then harvested by centrifugation at 10000 ×g for 15 min and washed in phosphate buffer (pH 7.0) twice. The bacterial pellet was washed in sterile water, re-centrifuged, and finally re-suspended in 5 ml sterile water. Artificially contaminated soil (5 g) in the centrifuge tubes was inoculated with small aliquots (up to 5 ml) of the final washed bacterial culture or 5 ml extract of button mushroom compost. After taking the weight of the tubes, they were wrapped with brown paper and placed on an orbital shaker at 200 rpm at 25°C. At the end of the period of 10 d, the weight of the tubes was recorded and 10 ml of sterile water were added to each tube to extract the soil water soluble heavy metals. The extracts were centrifuged at 10000 ×g for 10 min and filtered and the metal contents in the filtrate were determined using an ICP. Artificially contaminated soil without inoculation with the strain served as the control after centrifugation.

Results

Isolation and identification of heavy metal alleviating bacteria A total of 16 bacterial strains with the potential ability to alleviate heavy metals were screened based on estimating the metal concentration alleviated by each bacterium in medium supplemented with 3 different heavy metals (Co, Pb, and Cd). A bacterial strain showing the highest degree of metal resistance and metal alleviation was finally selected for the study. According to 16S rRNA sequence analysis, the selected strain showed close proximity with *Shewanella xiamenensis* FJ589031. Phylogenetic tree (Fig. 1) shows the position of the isolated bacterium with respect to related species which is a facultative anaerobe with the ability to reduce cadmium as well as iron and manganese metabolically; that is, it can use iron and manganese as the terminal electron acceptor in the electron transport chain (Ng *et al.* 2014).

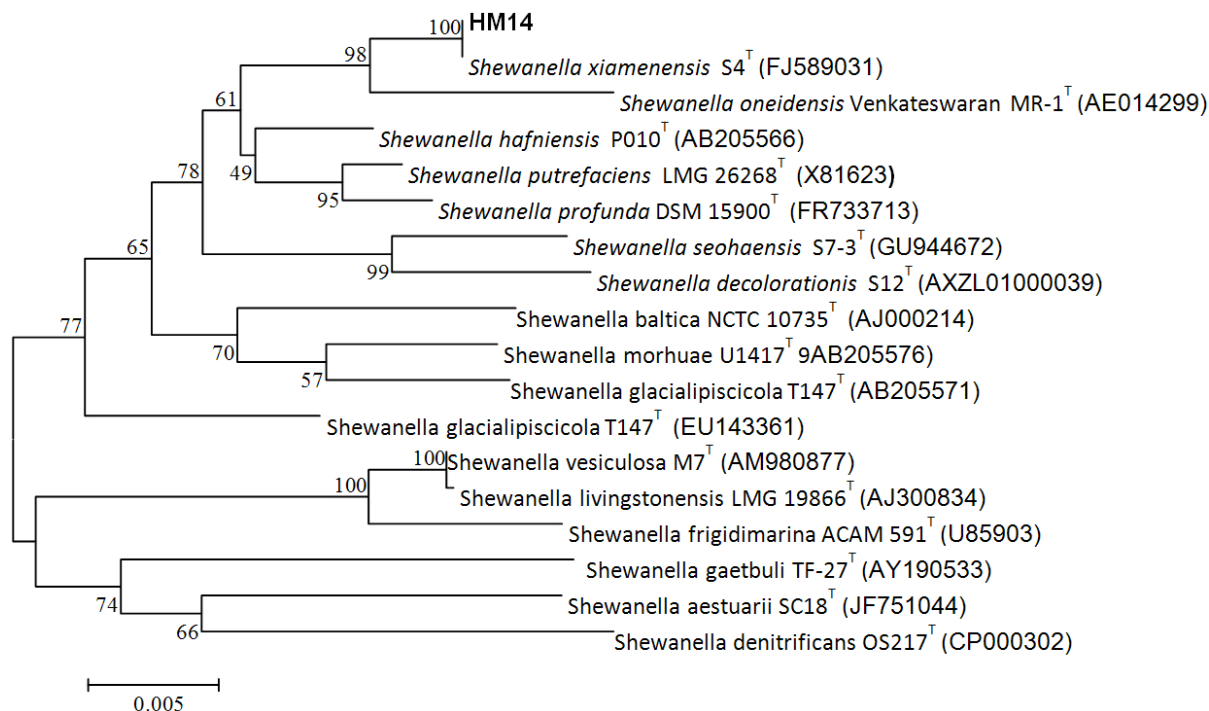


Fig. 1. Phylogenetic tree based on 16S rRNA gene sequences, showing the position of isolated metal solubilizing bacterial strain (HM14) with respect to related species. The scale bar indicates 0.002 substitutions per nucleotide position and accession numbers are given in parenthesis.

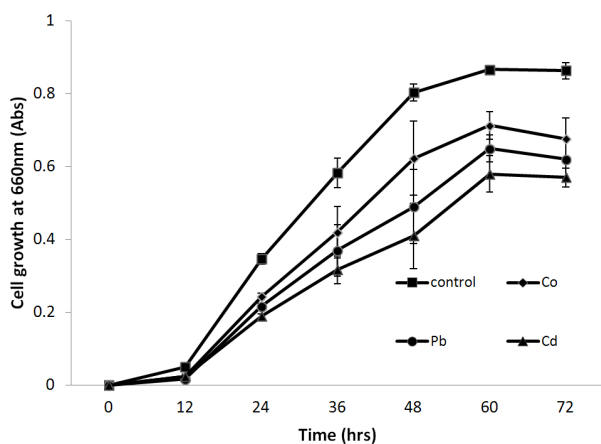


Fig. 2. Growth of *Shewanella xiamenensis* HM14 on sterilized liquid TSB medium supplemented with metals (Co, Pb and Cd) at the concentrations of 200 mg L⁻¹. Sterilized liquid TSB medium without supplemented with heavy metals was served as a control. Values are the means of three replicates. Error bars represent standard deviation.

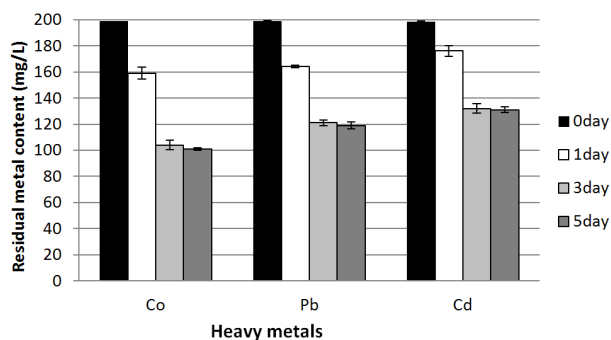


Fig. 3. Heavy metal alleviation by *Shewanella xiamenensis* HM14 on TSB medium supplemented with heavy metals (Co, Pb and Cd) at the concentration of 200 mg L⁻¹. Sterilized liquid TSB medium without supplemented with heavy metals was served as a control. Values are the means of three replicates. Error bars represent standard deviation.

Effect of heavy metals on bacterial growth The growth of the strain was monitored by measuring the optical density of culture supernatant supplemented with 3 different heavy metals (Co, Pb, and Cd) at definite time intervals. Results show that the bacterium is slowly growing in liquid media with 200 mg L⁻¹ of metals (Co, Pb and Cd) and the growth reaches to maximal at 60h after incubation. Although none of the metals was found to be highly toxic to the strain, slight

reductions in bacterial growth were observed in media supplemented media with Pb and Cd, compared to the metal free culture medium (Fig. 2). These results indicate that *S. xiamenensis* HM14 is able to maintain the resistance under high heavy metal-growth conditions.

Alleviation of heavy metals Alleviation effect of heavy metals was estimated by measuring the amount of metals

remaining in the TSB liquid medium inoculated with *S. xiamenensis* HM14. As depicted in Fig. 3, the strain was shown to be capable of reducing the amount of metal in the order of Co, Pb and Cd. Compared with the control, reductions of metals were 48, 39, and 34% for Co, Pb and Cd at 3day after incubation, respectively. This result represents that even growth of the strain somewhat resist at the presence of mg L⁻¹ of metals, high density of cell growth caused metal solubilization in medium.

Effect of bacterial strain on growth and metal uptake by *H. annuus* As showh in Table 1, inoculation with *S. xiamenensis* HM14 into sunflower (*Helianthus annuus*) pots resulted in increased fresh and dry biomass of *H. annuus* plants compared to non-inoculated plants. In case of the non-inoculated plants exposed by heavy metal stress, the growth of plant was inhibited in a significant level with $p < 0.05$. For instance, Cd toxicity caused 47 and 50% reductions in fresh and dry weight of the plant, respectively. Inoculation however led to increase in plant fresh and dry weight in the presence of heavy metals. The fresh and dry weight of the plants grown in Cd contaminated soils were respectively 18 and 15% higher than

those of non-inoculated plants. Similarly, in Pb contaminated soil, the percent increments were recorded as 20 and 15% respectively, and in Co contaminated soil, the corresponding figures were 20 and 28%.

The amounts of Co, Pb and Cd accumulated in the roots and shoots of *H. annuus* grown under inoculated and non-inoculated conditions are given in Table 2. Inoculation with *S. xiamenensis* HM14 resulted in increased accumulation of metals both in the shoots and roots. The accumulations of Co, Pb and Cd in shoots were respectively 28, 14 and 31% higher than those of non-inoculated plants. The corresponding accumulations for Co, Pb and Cd in roots were 17, 20 and 15% higher than those of non-inoculated plants. Regardless of inoculation or non-inoculation, the accumulation of metals in root system was found to be considerably higher than that of in shoots, which has been further confirmed by the low translocation factor (TF) for all the metals. However, TF of Co was somewhat higher than that of the other two metals. Similarly low bioconcentration factor (BCF) was also recorded from Pb and Cd. However, the results showed a good agreement and demonstrated that inoculation of the bacterial strain led to increase both TF and BCF of the three metals distinctly.

Table 1. Effect of inoculation with *Shewanella xiamenensis* HM14 on shoot and root weight of *Helianthus annuus*.

Metal	Treatment	Fresh weight (g/plant)		Dry weight (g/plant)	
		Shoot	Root	Shoot	Root
Metal free soil	control	1.74 (± 0.043)	0.109 (± 0.012)	0.089 (± 0.005)	0.036 (± 0.004)
	with strain	1.91 (± 0.035)	0.127 (± 0.008)	0.102 (± 0.004)	0.054 (± 0.007)
Co	control	1.32 (± 0.031)	0.062 (± 0.005)	0.073 (± 0.002)	0.021 (± 0.004)
	with strain	1.65 (± 0.042)	0.081 (± 0.007)	0.098 (± 0.003)	0.032 (± 0.002)
Pb	control	1.23 (± 0.027)	0.054 (± 0.003)	0.067 (± 0.001)*	0.023 (± 0.003)*
	with strain	1.51 (± 0.042)	0.085 (± 0.005)	0.075 (± 0.003)*	0.029 (± 0.004)*
Cd	control	1.08 (± 0.034)	0.047 (± 0.002)*	0.051 (± 0.003) *	0.015 (± 0.002) *
	with strain	1.30 (± 0.039)	0.054 (± 0.003)*	0.060 (± 0.004) *	0.018 (± 0.003) *

Values are means ($n=3$) ± standard deviation. Within each column, the means indexed by * are not significantly different at $p > 0.05$ between inoculated and non-inoculated plants according to Duncan's multiple range test.

Table 2. Effect of inoculation with *Shewanella xiamenensis* HM14 on accumulation and translocation of Co, Pb and Cd in *Helianthus annuus*.

Metal	Treatment	Metal content (mg/kg dry weight)		Bioconcentration Factor (BCF) ^a	Translocation Factor (TF) ^b
		Shoot	Root		
Co	control	19.24 (± 3.04)	81.23 (± 5.84)	0.456	0.237
	with strain	26.57 (± 3.57)	109.71 (± 6.27)	0.548	0.242
Pb	control	11.82 (± 0.87)*	68.53 (± 9.36)	0.343	0.172
	with strain	14.13 (± 0.91)*	85.26(± 7.37)	0.426	0.166
Cd	control	12.69 (± 6.43)	48.57 (± 6.25)*	0.243	0.261
	with strain	18.42 (± 12.25)	56.96 (± 7.24)*	0.285	0.323

^aBCF=metal concentration ratio of plant roots to soil; ^bTF=metal concentration ratio of plant shoots to roots.

Values are means ($n=3$) ± standard deviation. Within each column, the means indexed by * are not significantly different at $p > 0.05$ between inoculated and non-inoculated plants according to Duncan's multiple range test.

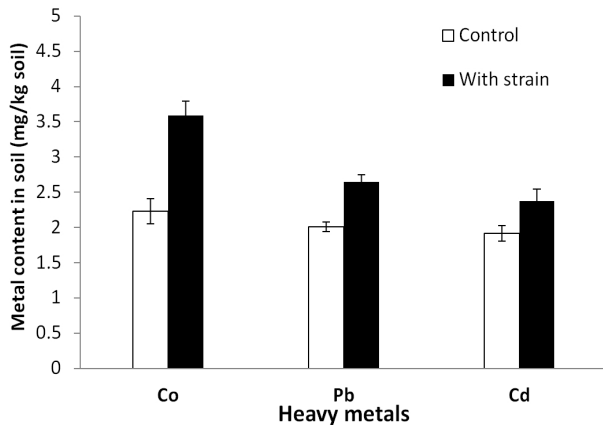


Fig. 4. Effect on mobilization of Co, Pb and Cd in soil by inoculation with *Shewanella xiamenensis* HMI4. Soil without inoculation the strain served as the control. Values are the means of three replicates. Error bars represent standard deviation.

Mobility of the metals in soil To investigate the impact of bacterial inoculation on the mobility of metals in soil, metal mobilization potential of the strain was assessed in a batch experiment with artificially contaminated soil. As given in Fig. 4, inoculation of the strain could increase the contents of water soluble metals from the soil extract, representing that heavy metals could be solubilized by microorganism inoculation. The mobilization of Co, Pb and Cd was respectively 38, 34 and 20% higher than those of the control soil.

Discussion

Growth response of the present strain under metal contamination conditions is in line with Rajkumar et al. (2008) and Prapagdee et al. (2013), who observed Zn, Cu and Ni resistance in *Bacillus weihenstephanensis* and Cd resistance in *Klebsiella* sp. BAM1. Generally microorganisms isolated from heavy metals contaminated soils possess the ability to withstand against multiple pollutants as they have adapted to such environments (Pal et al., 2005; Abou-Shanab et al., 2007).

The effectiveness of the strain as a plant growth-promoter was assessed with *Helianthus annuus*, a species known to have an ability to accumulate biomass rapidly and take up substantial amounts of metals (Turgut et al., 2004; Prapagdee et al., 2013). As reported by Ouzounidou et al. (2005) and El-Tayeb et al. (2006), accumulation of plant biomass could be affected by excessive concentrations of heavy metals, which exert adverse impacts on growth and function of root system resulting in poor uptake of water and nutrients. As reported by Jiang et al. (2008), inoculation with *Burkholderia* sp. J62 led to increase shoot and root dry weights of corn and tomato plants. Inoculation with *Pseudomonas fluorescens* PsIA12 resulted in enhanced growth of *Zea mays* and its

uptake of N, P and K (Egamberdiyeva et al., 2002). The content of P, K, S and Ca was reported to be increased by the inoculation of rhizobacteria in barley plant grown in metal contaminated soil (Belimov et al., 2004). According to them, inoculation with rhizobacteria resulted in 42% increase in growth of the barley plant compared to the control. Based on the results, they further stated that nutrients play an important role in the detoxification of heavy metals. Their findings were in line with Lebeau et al. (2008), who reported that rhizobacteria could have strong impacts on the nutritional status and the plant resistance to heavy metals. Most recently, Prapagdee et al. (2013) reported that growth of *H. annuus* could be enhanced by the inoculation of *Micrococcus* sp. MU1 and *Klebsiella* sp. BAM1 under Cd contaminated conditions. Belimov et al. (2001) also observed bacterial-assisted growth enhancement in *Brassica napus* grown in a soil contaminated with Cd. The plant growth-promoting potential of the present strain could be attributed at least partly to the phosphate solubilization ability of the strain under metal stress conditions. In this regards, Rajkumar et al. (2005) also reported that phosphate solubilization ability of *Pseudomonas* sp. could be contributed to the growth enhancement of the inoculated plants. Inoculation of phosphate solubilizing *Bacillus subtilis* SJ-101 resulted in higher shoot and root length and biomass with or without Ni (Zaidi et al., 2006). Bacteria is reported to promote the growth of plants (i) indirectly through producing antibiotics to inhibit soil pathogens, and (ii) directly through increasing nutrient and water uptake and thereby the plant biomass (Belimov et al., 2004). Through the production of siderophores, specific enzymes, and organic acids involved in phosphorus solubilization, and fixation of atmospheric N₂, bacteria could assist plants to withstand against metal toxicity (Kloepper, 2003). In this regards, Borgmann (2000) reported that *Kluyvera ascorbata* SUD165 protected *Brassica juncea* and *Brassica campestris* against Ni, Pb and Zn toxicity through the production of enzyme ACC deaminase. Plant growth promoting rhizobacteria was reported to enhance root elongation of *Brassica napus* by stimulating IAA synthesis (Sheng and Xia, 2006). In *Brassica juncea*, root elongation was reported to be enhanced by non-identified rhizobacteria (Belimov et al., 2005), *Variovorax paradoxus* 5C-2 (Belimov et al., 2005) and root dry weight was increased by rhizobacteria (Sheng and Xia, 2006).

Regardless of inoculation or non-inoculation, the accumulation of metals in root systems was found to be considerably higher than that of in shoots. This could primarily be attributed to the poor translocation of heavy metals from roots to shoots (Rajkumar et al., 2006). However, as shown in Table 2, translocation factor of the each metal was increased with the inoculation of the strain, which was of enormous practical significance. Furthermore, metal accumulations in both shoots and roots were found to be higher in inoculated plants

than those of non-inoculated plants. Similar observations were made by Rajkumar *et al.* (2008) for Zn accumulation in *H. annuus* inoculated with *Bacillus weihenstephanensis*. However, according to Wani *et al.* (2007), inoculation of *Bradyrhizobium* sp. on surface sterilized seeds of *Vigna radiate* reduced the concentration of Ni in roots, shoots and grains by 15, 19 and 22%, respectively, compared with non-inoculated plants.

The inter-relationships among soil pH, solubility and speciation of metals have been intensively investigated (Gadd, 2004). Bacteria such as *Azotobacter chroococcum* (N-fixing bacteria), *Bacillus megaterium* (P-solubilizer) and *Bacillus mucilaginosus* (K-solubilizer) (Wu *et al.*, 2006) and *Bacillus* sp. RJ16 (Sheng and Xia, 2006) were reported to decrease the pH, enhancing the bioavailability of Cd, Pb and Zn (Chen *et al.*, 2005). As stated by Zaidi *et al.* (2006), reduction in pH from 7.5 to 4.8 with the inoculation of phosphate solubilizing *Bacillus subtilis* SJ-101 possibly created favourable conditions for the solubilization of metals and their subsequent uptake by the plants. The increased accumulation of metals in the presence of bacterial strain might be due to the increased uptake of metals under acidic soil conditions created by the phosphate solubilization (Rajkumar *et al.*, 2008). Inoculation of Cd-resistant bacterial strains to *Brassica napus* to a metal contaminated soil significantly increased the plant uptake of Cd when compared with the non-inoculated controls, as a result of pH reduction (Sheng and Xia, 2006). However, on the contrary, *Glomus caledonium* (Chen *et al.*, 2004) and *Glomus mosseae* (Citterio *et al.*, 2005) were reported to have no effect on the speciation of Cd and Zn, and Cr and Ni, thus no effect of bioaugmentation by these arbuscular mycorrhizal (AM) fungi on the rate of phytoextraction has been observed, which could be attributed to whether strong symbiotic relationships between AM fungi and host plants.

The present findings of metal mobilization are in agreement with Wu *et al.* (2006) and Prapagdee *et al.* (2012) who also reported bacteria-assisted increase in heavy metal mobilization. Generally, the low amount of metals extracted by plants from a soil is attributed mainly to the low availability of metals. As reported by several authors, the available metal content in a soil is less than 1% of the total metal content (Whiting *et al.*, 2001; Braud *et al.*, 2006). Metal availability is influenced by the nature of the metal and soil characteristics such as pH, CEC and organic matter (Kayser *et al.*, 2001; Lebeau *et al.*, 2008). Bioaugmentation could enhance metal bioavailability by increasing the concentration of the available fractions. As revealed by the present results, the release of heavy metals from the non-soluble phases to soluble phases could be facilitated by the bacterial strain. Therefore, increased accumulation of metals, in particular Co in both the shoots and roots of *H. annuus* could be attributed to the higher water soluble metal contents in soil inoculated with bacterial strain. As reported by the results of previous studies, *H. annuus* is capable of

accumulating high amounts of Pb, Cd, Cu, Zn and Co, in both the shoots and the roots (Boonyapookana *et al.*, 2005; Marchiol *et al.*, 2007). According to Braud *et al.* (2006), inoculation of *Pseudomonas aeruginosa* and *Pseudomonas fluorescens* has resulted in 113% increment of Pb content in the exchangeable fraction of the soil. However, the Pb concentration bound to free Mn oxides, organic matter and in the residual fraction remained stable. Abou-Shanab *et al.* (2006) observed an increase of extractable Ni with *Microbacterium arabinogalactanolyticum* by a factor up to 15. As reported by Baum *et al.* (2006), the concentrations in NH₄NO₃-extractable Cd, Cu, Pb and Zn in a soil bioaugmented with ectomycorrhizal fungus *Paxillus involutus*, were 1.22-, 1.11-, 1.33- and 1.33-fold higher than those of non-bioaugmented soil, depending on the soil composition. However, comparing and contrasting of the results of bioaugmentation studies are hard to perform, because the estimation of bioavailable fraction of metals has been done under different conditions with different extractants such as water (Di Gregorio *et al.*, 2006; Wu *et al.*, 2006), MgCl₂ (Braud *et al.*, 2006), NH₄NO₃ (Baum *et al.*, 2006), NH₄O-Ac (Wu *et al.*, 2006), DTPA (Di Gregorio *et al.*, 2006; Wu *et al.*, 2006), KNO₃ (Di Gregorio *et al.*, 2006) and HCl (Wang *et al.*, 2007).

Conclusion

The strain, *Shewanella xiamenensis* HM14 was found to be capable of solubilizing metals (Co, Pb and Cd). Metal mobilization potential of the strain showed that inoculation could increase the concentrations of water soluble Co, Pb and Cd than those of non-inoculated soils. Inoculation with the strain also resulted in increased shoot and root biomass and enhanced accumulation of Co, Pb and Cd in *Helianthus annuus* plants. The strain was found to be capable of promoting metal translocation from the roots to the shoots of *H. annuus*. Therefore, *Shewanella xiamenensis* HM14 could be identified as an effective promoter of phytoextraction of Co, Pb and Cd from metal-contaminated soils.

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