#### Effects of Organotin Compounds on Follicular Steroidogenesis in Frogs

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ABSTRACT : Some organotin compounds such as butyltins and phenyltins are known to induce impo-sex in various marine animals and are considered to be endocrine disruptors. In this study, the effect of organotins on follicular steroidogenesis in amphibians was examined using ovarian follicles of Rana dybowskii and Rana catesbeiana. Isolated follicles were cultured for 6 or 18 h in the presence and absence of frog pituitary homogenate (FPH) or various steroid precursors, and the levels of product steroids in the culture media oassay. Among the butyltin compounds, tributyltin (TBT) strongly and dose-dependently inhibited the FPH-induced synthesis of pregnenolone  $(P_5)$  and progesterone  $(P_4)$  by the follicles. TBT also strongly suppressed the conversion of cholesterol to  $P_5$  and partially suppressed the conversion of  $P_5$  to  $P_4$ . A high concentration of dibutyltin (DBT) also inhibited steroidogenesis by the follicles while monobutyltin and tetrabutyltin had negligible effects. The toxic effect of TBT or DBT was irreversible and a short time of exposure (30 min) was enough to suppress steroidogenesis. All the phenyltin compounds significantly inhibited FPH-induced P<sub>5</sub> synthesis by the follicles. The effective dose of 50% inhibition by diphenyltin was 0.04  $\mu$ M and those of monophenyltin and triphenyltin were 0.24  $\mu$ M and 0.3  $\mu$ M, respectively. However, none of the phenyltin compounds significantly suppressed the conversion of P<sub>4</sub> to  $17 \alpha$ -hydroxyprogesterone (17  $\alpha$ -OHP) (by 17  $\alpha$ -hydroxylase), 17  $\alpha$ -OHP to androstenedione (AD) (by  $C_{17-20}$  lyase), or AD to testosterone by the follicles. Taken together, the data show that among the steroidogenic enzymes, P450scc in the follicles is the most sensitive to organotin compounds and that an amphibian follicle culture system can be a useful screening model for endocrine disruptors.

Key words : Amphibian, Endocrine disruptors, Steroidogenesis, Organotin compounds.

#### **INTRODUCTION**

The use of tributyltin (TBT) in antifouling paints and wood preservatives for ships is responsible for world-wide contamination of fresh water and marine environments. It is well known that these compounds, when released into the environment, can cause significant declines in the populations of marine oyster species (Laughlin & Linden, 1987). In particular, TBT is known to be highly toxic toward various non-target marine organisms (Fent, 1996). Phenyltin compounds are widely used in agriculture as fungicides and in consumer products (Kannan & Lee, 1996). They contaminate the environment and have toxic effects on aquatic and terrestrial animals including humans (Kimbrough, 1976; Kannan et al., 1997; Maguire, 2000; Nielsen & Strand, 2002). In particular, some studies have shown that they are responsible for the presence of imposex in a population of gastropods (Horiguchi et al., 1998; Pellizzato et al., 2004). Also, derivatives of these compounds are known to cause development and reproductive toxicity in

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mammals by disrupting the endocrine system. Recently, considerable amounts of TBT and triphenyltin (TPT) were found to have accumulated in starfish and bivalves collected from the coastal area of the Korean peninsula (Shim et al., 2005). Many reports have also indicated that organotins suppress various steroidogenic enzymes in different animal tissues. For example, TBT was found to inhibit the activities of  $3\beta$ -hydroxysteroid dehydrogenase ( $3\beta$ -HSD) and  $17 \alpha$ -hydroxylase, but not that of  $17 \beta$ -hydroxysteroid dehydrogenase (17 $\beta$ -HSD) in rat testes microsomes (McVey & Cooke, 2003). The toxic effects of butyltin compounds on steroidogenic enzymes have also been examined in cows (Schoenfelder et al., 2003; Yamazaki et al., 2005), pigs (Ohno et al., 2005) and humans (Cooke, 2002). Our previous studies also have shown that some butyltins suppressed follicular steroidogenesis in the frog; R. catesbeiana (Ahn et al., 2007).

The aim of this research is to determine whether phenyltins also suppress follicular steroidogenesis in amphibians and to compare the responsiveness of follicles obtained from two different species of frogs to organotins. The present data showed that phenyltins like butyltins are very effective in suppressing follicular steroidogenesis in amphibians and that frog follicles obtained from two different species of frogs exhibit similar responses to organotins. This finding supports the idea that an amphibian follicle culture system can be a useful screening model for organic endocrine disruptors.

#### MATERIALS AND METHODS

#### 1. Animals

Mountain frogs (*Rana dybowskii*) were collected in October-December from streams in the Chonnam area, in the southwestern part of the Korean peninsula, and bull-frogs (*R. catesbeiana*) were collected from fields in the Chonbuk area, in the western part of the Korean peninsula. The animals were kept in stainless steel boxes containing water and stones to simulate their natural habitat. They

were acclimatized to these conditions for several days and after one week they were used for experimentation.

#### 2. Hormones and Reagents

Steroids (25-OH-cholesterol, pregnenolone, P5; progesterone, P<sub>4</sub>) were purchased from Sigma-Aldrich (St. Louis, Mo). All the organotin compounds, monobutyltin (MBT), dibutyltin (DBT), TBT, tetrabutyltin (TeBT), monophenyltin (MPT), diphenyltin (DPT), and triphenyltin (TPT) were also purchased from Sigma-Aldrich (St. Louis, Mo). Steroid precursors were dissolved in a vehicle composed of ethanol and propylene glycol (1:1). The final concentration of the vehicle in the medium was below 0.5% v/v and was found not to influence steroid secretion (data not shown). All organotins were dissolved in dimethylsulfoxide (DMSO). Different concentrations of reagents were prepared by dilution of the stock solution with DMSO. Frog pituitary homogenate (FPH) was prepared as previously described (Kwon et al., 1993; Kwon & Ahn, 1994) and used to stimulate follicular steroidogenesis.

#### 3. Follicle Culture

After animals were killed by decapitation, ovaries were removed immediately and divided into fragments in amphibian Ringer's solution (AR). Follicles were manually isolated under magnification from other ovarian tissue using watchmaker's forceps under a dissecting microscope. Isolated follicles from R. dybowskii were cultured for 6 h and those from R. catesbeiana for 16 h AR in the presence or absence of FPH (0.1 pituitary/2 ml) and/or with various exogenous substrates. In vitro culture was carried out at room temperature (23-25°C, with 10 follicles in 1 ml of AR per well [24 wells per dish; Nunc, Denmark]). The duration of follicle culture and the dose of FPH were chosen on the basis of our previous data from Rana (Kwon & Schuetz, 1985; Kwon et al., 1991). Control and experimental groups were cultured in triplicate using follicles from each animal and each experiment was replicated with follicles from 3-6 frogs.

## 4. Follicle Extraction and Sample Preparation for RIA

As  $P_5$  was not secreted into the culture medium,  $P_5$  was extracted from follicles by a procedure previously described by Kwon and Schuetz (1985). After culture, follicles were extracted in the culture wells using methanol (GR, Merck, 1 ml methanol per 10 follicles per well) and shaken for 15 min (100 oscillations per min). Methanol extracts were dried under vacuum using a centrifugal lyophilizer (Savent SVC-100; Hickville, USA), and reconstituted using 0.05 M gelatin phosphate buffered saline (GPBS, pH 7.2). Extracts in GPBS were assayed without further purification. For other steroids, the culture medium was used directly for measuring the steroids by RIA.

#### 5. Steroid Radioimmunoassay

Steroids secreted by the ovarian follicle into the medium and those present in the follicle extracts, were measured by radioimmunoassay (RIA). The general assay procedure was adapted from those described in previous studies (Kwon et al., 1993; Kwon & Ahn, 1994). Labeled P<sub>5</sub> ([7-<sup>3</sup>H (N)]-pregnenolone; 25 Ci/mmol) was purchased from PerkinElmer Life Science (Boston, MA, USA). Labeled P<sub>4</sub> (1,2,6,7-<sup>3</sup>H-progesterone; 99 Ci/mmol),  $17 \alpha$ -OHP ([1,2,6,7,-<sup>3</sup>H])-hydroxyprogesterone; 58.5 Ci/mmol), labeled AD ([1,2,6,7-<sup>3</sup>H]-androstenedione; 86.1 Ci/mmol) and T ([1,2,6,7-3H]-testosterone; 98 Ci/mmol) were obtained from Amersham (Buckinghamshire, England). The steroid antiserum against AD was purchased from Sigma (St. Louis, MO, USA) and those against  $P_5$ ,  $P_4$ , 17  $\alpha$  -OHP, and testosterone (T) were obtained from Biogenesis (England). The P<sub>5</sub> antiserum cross-reacted 19% with P<sub>4</sub>, and less than 3% with 17  $\alpha$ -OHP, AD, or T. The P<sub>4</sub> antiserum crossreacted 4% with T, 2% with 17  $\alpha$  -OHP and less than 0.1% with other steroids. The 17  $\alpha$  -OHP antiserum cross-reacted 0.25% with P<sub>4</sub>, and less than 0.1% with other steroids. The AD antiserum cross-reacted 4.5% with T, and less than 0.1% with other steroids. The T antiserum cross-reacted 3.3% with 11  $\beta$ -hydroxy testosterone, and less than 0.1% with E<sub>2</sub>, P<sub>4</sub> or other steroids. Each sample was quantified for tritium using a Packard Tri-Carb 2900TR liquid scintillation analyzer. Routinely, duplicate steroid standards were included in each assay (5-2,000 pg). The between and within assay coefficients of variation (CV) for P<sub>5</sub> were 9.2% and 9.3%, respectively. The CVs for P<sub>4</sub> were 11.5% and 6.7%, 6.2% and 6.5% for 17  $\alpha$ -OHP, 5.6% and 6.9% for AD, 6.3% and 7.9% for T, respectively.

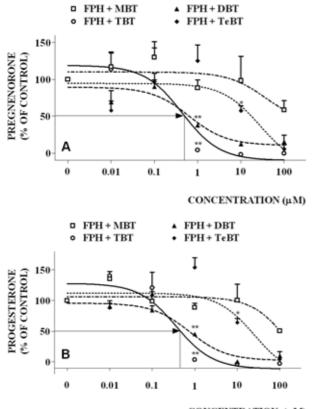
#### 6. Statistical Analysis

Data (% of control) for all experiments were transformed using an arcsine-square root transformation and were analyzed by one-way analysis of variance (ANOVA) or Student's *t*-test using Prism statistical software (Kwon et al., 1993; Kwon & Ahn, 1994). Differences were considered to be significant when their probability was less than 0.05 or 0.01.

#### RESULTS

1. Effect of Butyltins on  $P_5$  and  $P_4$  Synthesis by Ovarian Follicles of *R. dybowskii In Vitro* 

In previous studies, we showed that butyltin compounds suppress follicular steroidogenesis of *R. catesbeiana* (Ahn et al., 2007). In the present experiments, we performed similar experiment with the ovarian follicles of the Korean frog, *R. dybowskii* in order to know whether the follicles from the two species respond similarly to butyltin compounds. The isolated follicles were incubated for 6 h in AR in the presence of FPH and/or various concentrations of butyltins (0.01-100  $\mu$ M) and P<sub>5</sub> in follicles or P<sub>4</sub> in culture media were measured by RIA. Fig. 1 shows that TBT strongly suppressed FPH-induced P<sub>5</sub> and P<sub>4</sub> synthesis by the follicles in a dose-dependent manner (ED<sub>50</sub>, 0.7  $\mu$ M) (Fig. 1). DBT also inhibited P<sub>5</sub> and P<sub>4</sub> synthesis in a similar way but MBT and TeBT were much less inhibitory (Fig. 1).

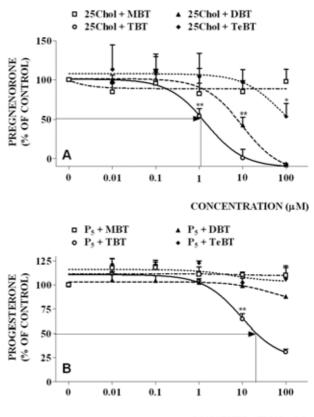


CONCENTRATION (µM)

Fig. 1. Effects of butyltins on FPH-induced P<sub>5</sub> and P<sub>4</sub> synthesis by ovarian follicles of *Rana dybowskii*. Isolated follicles were cultured for 6 h in the presence of FPH (0.05 gland/ mℓ) with or without various concentrations of butyltins. Each point in the figure represents relative P<sub>5</sub> (A) or P<sub>4</sub> (B) levels (mean ± SEM, % of control) (n=6, duplicate incubation, 3 animals). \* p<0.05, \*\* p<0.01 when compared to control by *t*-test.

### 2. Effect of Butyltins on the Conversion of 25-OH-Cholesterol to $P_5$ and $P_5$ to $P_4$ by Ovarian Follicles of *R. dybowskii*

The effect of butyltin compounds on the conversion of cholesterol to  $P_5$  by P450scc was examined. As it was known that exogenous cholesterol could not be utilized by the follicles, the steroid precursor 25-OH-cholesterol (25 Chol) was used as precursor for  $P_5$  synthesis. Isolated follicles were cultured for 6 h in the presence of 25-OH-cholesterol (100 ng/m $\ell$ ) with or without various con-



CONCENTRATION (µM)

Fig. 2. Effects of butyltins on the conversion of 25-OH-cholesterol to  $P_5$  and  $P_5$  to  $P_4$  by *R. dybowskii* follicles. Isolated follicles were cultured for 6 h in the presence of 25-OH-Cholesterol (100 ng) or  $P_5$  (100 ng) with or without various concentration of butyltin. Each point in panel A represents relative  $P_5$  levels (mean  $\pm$  SEM, % of control) and panel B represents relative  $P_4$  levels (n=6 duplicate incubations, 3 animals).

centrations of butyltins; after culture,  $P_5$  levels in the follicles were measured. Fig. 2 shows that TBT strongly suppressed the conversion of 25-OH-cholesterol to  $P_5$  in a dose-dependent manner (ED<sub>50</sub>, 1  $\mu$ M). DBT also suppressed the conversion in high doses (>10  $\mu$ M), but MBT and TeBT failed to suppress the conversion (Fig. 2A). The effect of butyltins on the conversion of  $P_5$  to  $P_4$  by the follicles was also examined (Fig. 2B). Interestingly, only TBT suppressed the conversion of  $P_5$  to  $P_4$  at relatively high doses (ED<sub>50</sub>, 13  $\mu$ M), and other butyltins failed to suppress the conversion even at the highest doses of butyltins (100  $\mu$  M). Thus it is evident that TBT and DBT suppressed the enzyme P450scc enzyme that is responsible for the conversion of cholesterol to P<sub>5</sub> and were much less effective inhibitors of the enzyme 3 $\beta$ -HSD that is responsible for the conversion of P<sub>5</sub> to P<sub>4</sub>.

3. Irreversible Inhibition by TBT and DBT of Follicular Steroidogenesis in *R. dybowskii* 

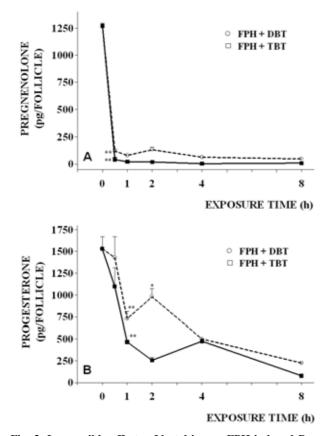
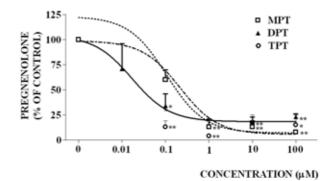


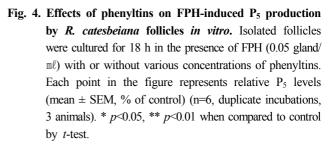
Fig. 3. Irreversible effects of butyltins on FPH-induced  $P_5$  or  $P_4$  production by *R. dybowskii* follicles. Isolated ovarian follicles were exposed to 10  $\mu$  M butyltins (DBT and TBT) for designated time periods (0.5, 1, 2, 4 and 8 hours). After washing, the follicles were cultured further for 18 h in the presence of FPH. After culture,  $P_5$  in the follicles (panel A) or  $P_4$  in the medium (panel B) were measured. Each point in the figure represents relative  $P_5$  or  $P_4$  levels (mean  $\pm$  SEM, % of control) (n=6, duplicate incubations, 3 animals).

In order to know whether the toxic effect of the butyltins is reversible, we exposed isolated follicles to high doses of TBT or DBT (10  $\mu$  M) for various periods (0.5-8 h) and then transferred them to plain medium for further culture in the presence of FPH for 6 h. After culture, P<sub>5</sub> in the follicles and P<sub>4</sub> in the medium were measured. A half-hour exposure to the butyltins completely suppressed FPH-stimulated P<sub>5</sub> synthesis by the follicles (Fig. 3A). P<sub>4</sub> production by the follicles was also suppressed when they are exposed for 1 h or longer (Fig. 3B). Thus it is evident that the toxic effect of butyltins is irreversible.

# 4. Effects of Phenyltins on FPH-induced Pregnenolone (P<sub>5</sub>) Synthesis by the Ovarian Follicles of *R. catesbeiana*

The effect of phenyltins on the FPH-stimulated P<sub>5</sub> synthesis by the follicles was examined. The activity of P450scc can be estimated from the amount of P<sub>5</sub> produced by the follicles in response to phenyltins. Isolated follicles were cultured for 16 h in the presence of FPH (0.05 gland/mℓ) and/or phenyltin compounds of various concentrations (0.01-100  $\mu$  M). After culture, the amount of P<sub>5</sub> in the follicles was measured by RIA. A considerable amount of P<sub>5</sub> was





produced in response to FPH by the follicles (658-1,058 pg/follicle, n=6). As shown in Fig. 4, all the phenyltins strongly suppressed FPH-induced P<sub>5</sub> production by the follicles in a dose-dependent manner (Fig. 4). DPT was most effective in the inhibition (ED<sub>50</sub>, 0.04  $\mu$  M) and almost completely suppressed P<sub>5</sub> synthesis from very low concentration (0.01  $\mu$  M). MPT and TPT also inhibited the steroid synthesis significantly from 0.1  $\mu$  M (ED<sub>50</sub>, 0.24  $\mu$  M and 0.3  $\mu$  M, respectively). Therefore, the toxicities of phenyltin compounds for P450scc were in the order DPT > MPT = TPT.

5. Effects of Phenyltins on the Conversion of P<sub>5</sub> to P<sub>4</sub> or P<sub>4</sub> to 17  $\alpha$ -OHP by Ovarian Follicles of *R. catesbeiana* 

To assess the effect of phenyltin compounds on  $3\beta$ -hydroxysteroid dehydrogenase (3  $\beta$  -HSD), which converts P<sub>5</sub> to P<sub>4</sub>, or 17  $\alpha$  -hydroxylase, which converts P<sub>4</sub> to 17  $\alpha$  - OHP, in the follicles, we cultured ovarian follicles for 16 h in medium containing P<sub>5</sub> (100 ng/m $\ell$ ) or P<sub>4</sub> (100 ng/m $\ell$ ); after culture, the amount of P<sub>4</sub> converted from P<sub>5</sub> or  $17 \alpha$  - OHP converted from P4 was measured in the medium. As shown in Fig. 5A, all the phenyltins seemed to suppress the conversion of  $P_5$  to  $P_4$  at high doses (>10  $\mu$  M). However, the P<sub>4</sub> levels converted from P<sub>5</sub> exhibited considerable variation and were not significantly suppressed by the phenyltins in the medium (Fig. 5A). Thus it seems that the phenyltins partially, but not significantly, suppressed  $3\beta$ -HSD activity in the follicles. Interestingly, the conversion of P<sub>4</sub> to  $17 \alpha$ -OHP by follicles was not affected by phenyltins in the medium even at the highest dose examined (100  $\mu$  M) (Fig. 5B). TPT instead slightly, but not significantly, increased the accumulation of  $17 \alpha$ -OHP. The accumulation of  $17 \alpha$ -OHP likely resulted from the inhibition of its metabolism rather than from stimulation of  $17 \alpha$ -hydroxylase by the TPT (Fig. 5B). These results indicated that  $17 \alpha$ -hydroxylase activity is not sensitive to phenyltin compounds.

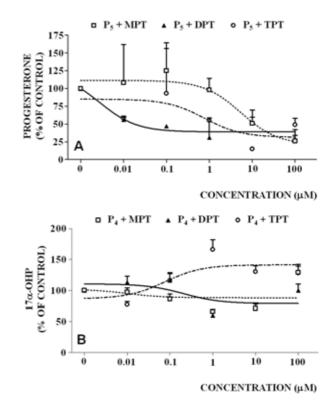


Fig. 5. Effects of phenyltins on the conversion of  $P_5$  to  $P_4$  or  $P_4$  to  $17 \alpha$ -OHP by ovarian follicles of *R. catesbeiana*. Isolated follicles were cultured for 18 h in the presence of  $P_5$  or  $P_4$  with or without various concentrations of phenyltins. After culture,  $P_4$  or  $17 \alpha$ -OHP in medium were measured by RIA. Each point in panel A represents relative  $P_4$  levels and panel B represents relative  $17 \alpha$ -OHP levels (mean  $\pm$  SEM, % of control) (n=6, duplicate incubations, 3 animals).

# 6. Effects of Phenyltin Compounds on the Conversion of $17 \alpha$ -OHP to AD and AD to T by the Ovarian Follicles of *R. catesbeiana*

To ascertain whether phenyltins have any effect on the activity of C<sub>17-20</sub> lyase which converts  $17 \alpha$ -OHP into androstenedione (AD) or  $17\beta$ -hydroxysteroid dehydrogenase ( $17\beta$ -HSD) which converts AD to T in the follicles, we cultured ovarian follicles in medium containing  $17 \alpha$ -OHP (100 ng/m $\ell$ ) or AD as precursors, and measured levels of product steroids (AD and T, respectively) in the medium were measured after 16 h of culture. The amount of AD

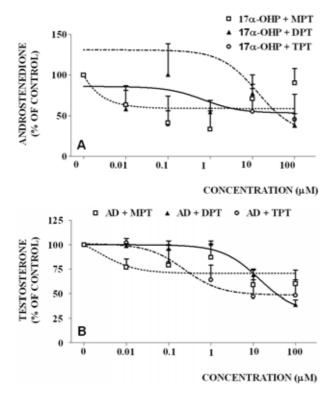


Fig. 6. Effects of phenyltins on the conversion of 17 a-OHP to AD or AD to T by ovarian follicles of *R. catesbeiana in vitro*. Isolated follicles were cultured for 18 h in the presence of 17 a-OHP or AD with or without various concentrations of phenyltins. After culture, AD or T in the medium were measured by RIA. Each point in panel A represents relative AD levels and panel B represents relative T levels (mean ± SEM, % of control) (n=6, duplicate incubations, 3 animals).

secreted by the control follicles was in the range of 11-235 ng/follicle (n=6). Without addition of 17  $\alpha$ -OHP, barely detectable levels of AD were produced. As shown in Fig. 6A, all the phenyltins (MPT, DPT, and TPT) partially, but not significantly, suppressed AD production by the follicles at relatively high doses (Fig. 6A). The levels of AD exhibited considerable variation. When follicles were cultured in medium containing AD (100 ng/ml), the amounts of T in the medium (control group) were in the range of 234-1,984 pg/follicle (n=6). As shown in Fig. 6B, all the phenyltins partially, but not significantly, suppressed the conversion of AD to T at high doses (> 10  $\mu$  M). Taken

together, these data showed that  $C_{17-20}$  lyase and  $17\beta$  - HSD in the follicles are not critically sensitive to the phenyltins.

#### DISCUSSION

In this study, we have demonstrated that some butyltins and phenyltins strongly suppressed FPH-induced P<sub>5</sub> and P<sub>4</sub> production by ovarian follicles of *R. dybowskii* and *R. catesbeiana in vitro*, probably through suppression of the P450scc enzyme. The other steroidogenic enzymes in the follicles such as  $3\beta$ -HSD, C<sub>17-20</sub> lyase and  $17\beta$ -HSD were much less sensitive to these organotins than P450scc and were not significantly suppressed by the organotins.  $17 \alpha$  - hydroxylase was not affected by the organotins even at high doses. The toxic potential of butyltins followed the order TBT>DBT>TeBT>MBT. However, all the phenyltins (MPT, DPT, and TPT) exhibited similar toxicity to the follicular enzymes; they strongly suppressed P450scc and partially suppressed the other enzymes.

Organotin compounds are important commercial organometals, used widely as agricultural biocides, marine antifouling paints and industrial catalysts. Environmental contamination resulting from extensive use of organotin compounds has been of great concern because of their deleterious effects on non-target organisms including humans (Kimbrough, 1976; Maguire, 2000; Nielsen & Strand, 2002). With their increased usage, considerable attention has focused on potential toxicity of these compounds. When TPT accumulates in tissues of fish, it has been shown to inhibit cytochrome P-450scc in fish hepatoma cells (Bruschweiler et al., 1996; Ohhira & Matsui, 1996; Schwaiger et al., 1996). Previous studies demonstrated that phenyltin compounds (1  $\mu$ M, *in vitro* exposure for 1 h) could cause considerable inhibition of the tumor-killing function of human NK cells. The toxic potential of the phenyltin compounds followed the order TPT> DPT> MPT (Whalen et al., 2000; Whalen et al., 2002). Human exposure to phenyltin compounds might come from occupational exposure and/or consumption of contaminated food (Tolosa et al., 1992; Kannan et al., 1995; Ueno et al., 1999; Whalen et al., 2000). Natural killer (NK) cell are a primary immune defense against tumor and virally infected cell. Thus, phenyltin compounds may pose a threat of increased viral infection or tumorigenesis if they interfere with NK-cell function.

Several studies have reported that butyltin compounds suppressed steroidogenic enzyme activities. TBT directly inhibits the activities of steroidogenic P450scc and  $3\beta$ -HSD at concentrations higher than 10  $\mu$ M in bovine adrenal cultured cells (Yamazaki et al., 2005). It has also been reported that TBT at 12  $\,\mu\,\mathrm{M}$  and DBT at 74  $\,\mu\,\mathrm{M}$ inhibited human aromatase activity, and that MBT and TeBT were not suppressive (Cooke, 2002). TBT is also a competitive inhibitor of aromatase activity, with an IC<sub>50</sub> value for 6.2  $\mu$  M in human placental aromatase (Heidrich et al., 2001). McVey and Cooke (2003) found TBT to be the most potent inhibitor among butyltins of  $3\beta$ -HSD activity in rat testis microsomes. These results indicate that butyltin compounds suppress steroidogenic enzyme activities in various mammalian cells, and that TBT is the most effective inhibitor. Previous study in our laboratory demonstrated that some kinds of butyltins (BTs) have inhibitory effects on the follicular steroidogenesis of R. catesbeiana (Ahn et al., 2007).

In this study, we have shown that the response of *R*. *dybowskii* ovarian follicles to organotin compounds is very similar to that of *R*. *catesbeiana*. Thus the data presented here and in our previous study (Ahn et al., 2007) demonstrate that some butyltins and phenyltins suppress steroid-ogenic enzymes, particularly P450scc, in *Rana* ovarian follicles *in vitro* as observed in fish and mammal. In general, the sensitivity of frog ovarian follicles cultured *in vitro* to butyltins appeared to be similar to or more sensitive than those obtained from various animal cell lines. For example, TBT suppressed conversion of cholesterol to pregnenolone (P450scc activity) at 1  $\mu$ M in *R*. follicles *in vitro* (Fig. 2) while TBT suppressed the P450scc activity

in bovine adrenal cultured cells at higher than 10  $\mu$ M (Yamazaki et al., 2005). Because the frog follicle culture system is simple and cheap to maintain, it may provide a practical screening model to evaluate the potential effects of various endocrine disrupte(o)rs or environmental pollutants.

Population declines in some species of amphibians have been documented worldwide but causes of these declines remain unclear. Amphibians are vulnerable to water contamination because they have aquatic life histories and highly permeable skin (Pickford & Morris, 1999). Our study demonstrated that amphibian follicles are more sensitive to organotins (endocrine disruptors) than ovarian or testis cells of fish or mammals and this may be one of the reasons for the decline of amphibian species in nature. Further studies are needed to establish the amphibian follicle culture model as an experimental model for screening endocrine disruptors.

#### ACKNOWLEDGEMENTS

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#### REFERENCES

- Ahn RS, Han SJ, Kim, SC, Kwon HB (2007) Effects of butyltin compounds on follicular steroidogenesis in the bullfrog (*Rana catesbeiana*). Environ Toxicol Pharmacol 24:149-154.
- Bruschweiler BJ, Wurgler FE, Fent K (1996) Inhibition of cytochrome P4501A by organotins in fish hepatoma cells PLHC-1. Environ Toxicol Chem 15:728-735.
- Cooke GM (2002) Effect of organotins on human aromatase activity *in vitro*. Toxicol Lett 126:121-130.
- Fent K (1996) Ecotoxicology of organotin compounds. Crit Rev Toxicol 26:1-117.
- Heidrich DD, Steckelbroeck S, Klingmuller D (2001) Inhibition of human cytochrome P450 aromatase activity by butyltins. Steroids 66:763-769.

- Horiguchi T, Hyeon-Seo C, Shiraishi H, Shibata Y, Soma M, Morita M, Shimizu M (1998) Field studies on imposex and organotin accumulation in the rock shell, *Thais clavigera*, from the Seto Inland Sea and the Sanriku region, Japan Sci Total Environ 214:65-70.
- Kannan K, Lee RF (1996) Triphenyltin and its degradation products in foliage and soils from sprayed pecan orchards and in fish from adjacent ponds. Environ Toxicol Chem 15:1492-1499.
- Kannan K, Tanabe S, Tatsukawa R (1995) Phenyltin residues in horseshoe crabs, *Tachypleus tridentatus* from Japanese coastal waters. Chemosphere 30:925-932.
- Kannan K, Senthilkumar K, Loganathan BG, Takahashi S, Odell DK, Tanabe S (1997) Elevated accumulation of tributyltin and its degradation products in bottle nose dolphins (*Tursiops truncates*) found stranded along the US Atlantic and Gulf coasts. Environ Sci Technol 31:296-301.
- Kimbrough RD (1976) Toxicity and health effects of selected organotin compounds: a review. Environ Health Perspect 14:51-56.
- Kwon HB, Schuetz AW (1985) Dichotomous effects of forskolin on somatic and germ cell components of the ovarian follicle: evidence of cAMP involvement in steroid production and action. J Exp Zool 236:219-228.
- Kwon HB, Ahn RS (1994) Relative roles of theca and granulosa cells in ovarian follicular steroidogenesis in the amphibian, *Rana nigromaculata*. Gen Comp Endocrinol 94:207-214.
- Kwon HB, Choi HH, Ahn RS, Yoon YD (1991) Steroid production by amphibian (*Rana nigromaculata*) ovarian follicles at different developmental stages. J Exp Zool 260:66-73.
- Kwon HB, Ahn RS, Lee WK, Im WB, Lee CC, Kim K (1993) Changes in the activities of steroidogenic enzymes during the development of ovarian follicles in *Rana nigromaculata*. Gen Comp Endocrinol 92:225-232.
- Laughlin RJ, Linden O (1987) Tributyltin contemporary environmental issues. Ambio Stockholm 26:252-256.

- Maguire RJ (2000) Review of the persistence, bioaccumulation and toxicity of tributyltin in aquatic environments in relation to Canada's toxic substances management policy. Water Quality Research Journal of Canada 35:633-679.
- McVey MJ, Cooke GM (2003) Inhibition of rat testis microsomal 3beta-hydroxysteroid dehydrogenase activity by tributyltin. J Steroid Biochem Mol Biol 86:99-105.
- Nielsen JB, Strand J (2002) Butyltin compounds in human liver. Environ Res 88:129-133.
- Ohhira S, Matsui H (1996) Comparative study of the metabolism of triphenyltin in hamsters and rats after a single oral treatment with triphenyltin chloride. Toxicol Lett 85:3-8.
- Ohno S, Nakajima Y, Nakajin S (2005) Triphenyltin and Tributyltin inhibit pig testicular 17beta-hydroxysteroid dehydrogenase activity and suppress testicular testosterone biosynthesis. Steroids 70:645-651.
- Pellizzato F, Centanni E, Marin MG, Moschino V, Pavoni B (2004) Concentrations of organotin compounds and imposex in the gastropod *Hexaplex trunculus* from the Lagoon of Venice. Sci Total Environ 332:89-100.
- Pickford DB, Morris ID (1999) Effects of endocrine-disrupting contaminants on amphibian oogenesis: methoxychlor inhibits progesterone-induced maturation of *Xenopus laevis* oocytes *in vitro*. Environ Health Perspect 107: 285-292.
- Schoenfelder M, Schams D, Einspanier R (2003) Steroidogenesis during *in vitro* maturation of bovine cumulus oocyte complexes and possible effects of tri-butyltin on granulosa cells. J Steroid Biochem Mol Biol 84: 291-300.
- Schwaiger J, Fent K, Stecher H, Ferling H, Negele RD (1996) Effects of sublethal concentrations of triphenyltinacetate on rainbow trout (*Oncorhynchus mykiss*). Arch Environ Contam Toxicol 30:327-334.
- Shim WJ, Yim UH, Kim NS, Hong SH, Oh JR, Jeon JK, Okamura H (2005) Accumulation of butyl- and phenyltin compounds in starfish and bivalves from the coastal

environment of Korea. Environ Pollut 133:489-499.

- Tolosa I, Merlini L, de Bertrand N, Bayona JM, Albaiges J (1992) Occurrence and fate of tributyl- and triphenyltin compoinds in western Mediterranean coastal enclosures. Environ Toxicol Chem 11:145-155.
- Ueno S, Susa N, Furukawa Y, Komatsu Y, Koyama S and Suzuki T (1999) Butyltin and phenyltin compounds in some marine fishery products on the Japanese market. Arch Environ Health 54:20-25.
- Whalen MM, Hariharan S, Loganathan BG (2000) Phenyltin inhibition of the cytotoxic function of human natural killer cells. Environ Res 84:162-169.

- Whalen MM, Green SA, Loganathan BG (2002) Brief butyltin exposure induces irreversible inhibition of the cytotoxic function on human natural killer cells, *in vitro*. Environ Res 88:19-29.
- Yamazaki T, Shimodaira M, Kuwahara H, Wakatsuki H, Horiuchi H, Matsuda H, Kominami S (2005) Tributyltin disturbs bovine adrenal steroidogenesis by two modes of action. Steroids 70:913-921.

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