Photocatalytic Decomposition of Gaseous Ozone over TiO₂ Thin Film

Ki-Chul Cho*, Kyung-Chul Hwang, Hyun-Gu Yeo¹, Taizo Sano², Koji Takeuchi²) and Sadao Matsuzawa²)

Department of Environmental Science, DongNam Health College, Suwon, 440–714, Korea

1)Department of Environmental Engineering, Konkuk University Seoul, Korea

2)National Institute of Advanced Industrial Science and Technology,
Institute for Environmental Management Technology, Photoenergy Application Group,

AIST Tsukuba West, Tsukuba 305–8569, Japan

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Abstract

The characteristics of heterogeneous photocatalytic decomposition were investigated at low concentration level of O_3 on TiO_2 for various operating parameters such as: loaded catalyst weight ($0 \sim 4 \text{ mg/cm}^2$), initial concentration of O_3 ($0.06 \sim 10.0$ ppm), gas flow rate ($1.0 \sim 2.5$ l/min), and relative humidity ($0 \sim 80\%$). This study was conducted using a flow-type reactor at room temperature. Three kinds of pure TiO_2 (P25, ST-01, and E-23) were employed as photocatalyts. It was found that O_3 removal ratio was identical, regardless of the loaded TiO_2 weight in the range from 0.5 to 4.0 mg/cm^2 . It was also found that higher initial ozone concentration results in greater oxidation rate of ozone and experimental data show kinetically a good agreement with Langmur-Hinshelwood kinetic model. We also observed that the removal ratio of O_3 decreases linearly with the increasing flow rate and also with the increasing relative humidity for each catalyst.

Key words: Photocatalyst, Langmur-Hinshelwood, Kinetic model, Flow-type reactor

1. INTRODUCTION

Ozone (O_3) has both beneficial and harmful role to human body. O_3 protects us from harmful UV radiation in the stratosphere but is a pollutant with high toxicity in the troposphere.

There is an increasing demand for utilization of O₃ as a markedly effective gaseous oxidizing agent: for example water purification, decolorization, and deodor-

Tel: +82-(0)31-249-6474, E-mail: ckc@dongnam.ac.kr

ization etc. For environmental protection, the decomposition of the remaining unreacted O_3 in the exhaust from the O_3 treatment apparatus is very important. Furthermore, lowering concentration of O_3 in indoor and outdoor air will be demanded in near future.

The decomposition of O_3 is an environmentally significant reaction and has attracted considerable attention. But, only a limited number of reports for the interaction of O_3 with solid surfaces are available (Oyama, 2000; Li *et al.*, 1998; Imamura *et al.*, 1991). Among many proposed processes for O_3 decomposition, adsorption and reaction with activated charcoal and catalytic

^{*} Corresponding author.

decomposition over metal and/or metal oxide catalysts are available methods for industrial application (Dhandapani and Oyama, 1995; Rakitskaya *et al.*, 1994; Calderbank and Lewis, 1976; Deitz and Bitner, 1973). But, for these methods, undesirable consumption of charcoal and additive energy is usually required.

Titanium dioxide (TiO₂) has been extensively used as a photocatalyst for various redox reactions as well as a catalyst for thermal reactions. Ohtani *et al.* (1992) studied the dark and photo O₃ decomposition at room temperature over 80 ppm of initial concentration of ozone and clarified the fundamental chemical feature of the photo-reaction. Bulanin *et al.* (1995) studied the low-temperature O₃ adsorption on TiO₂ using infrared (IR) spectroscopy. Pichat *et al.* (2000) investigated the decomposition of O₃ (50~2000 ppm) on TiO₂ in the temperature range of $248 \sim 358$ K.

No results, however, for the decomposition of O_3 of the ambient air concentration level (<1 ppm) on TiO_2 photocatalyst or any influences of environmental factors such as humidity, flow rate, and temperature etc. on the reaction have been published. The purpose of the present work is to investigate the characteristics of heterogeneous photocatalytic decomposition of low concentration level of O_3 on TiO_2 . The effect of operating parameters such as loaded catalyst weight, initial concentration of O_3 , gas flow rate, and water vapor concentration was evaluated at room temperature on various TiO_2 catalysts.

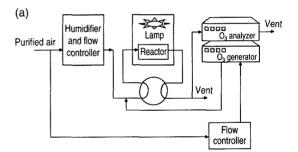
2. EXPERIMENTAL

Three kinds of TiO_2 were employed: P25 (Nippon Aerosil Ltd., 80% anatase, crystallite size of 25 nm, specific surface area of 50 m²/g), ST-01 (Ishihara Sangyo Ltd., 100 % anatase, crystallite size of 7 nm, specific surface area of 300 m²/g), and E-23 (Ecodevice Ltd. 100% anatase, crystallite size of 100 nm, specific surface area of 140 m²/g).

Figure 1(a) shows a schematic flow diagram of the reaction system. Decomposition of O₃ was conducted

in the flow-type photochemical reactor (Figure 1(b)) designed and fabricated in our laboratory (5 cm in width, 30 cm in length and 0.5 cm in height). Three 10 W black light lamps ($300 \sim 400$ nm, FL10NBL, Toshiba Ltd.) were used as a light source. The area of glass plate is 50 cm^2 (5 cm width and 10 cm in length), and the TiO₂ powder of 0.05 g was coated on it. Namely, a paste-like mixture of TiO₂ (0.05 g) and distilled water (3 ml) was applied onto the surface of glass plate (1 mg/cm²) and dried overnight at 383 K.

O₃ decomposition was carried out at room temperature (293~298 K) and atmospheric pressure. For analysis of O₃, a system (ML 9811, Monitor Labs, Inc., U.S.A.) composed of O₃ generator and analyzer was used. The measurement method is UV absorption using a single glass measurement cell and is based on the Beer-Lambert relationship to calculate the ozone concentration. O₃ concentration was continuously monitored every 10 seconds. In the experiment, air was purified using air purification apparatus (SGPU-22, STEC Ltd., JAPAN). To prepare humidified air, the purified



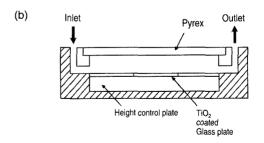


Fig. 1. Schematic depiction of the flow-type photochemical reaction system. (a) flow diagram, (b) detailed picture of reactor.

air was flown through purified water.

3. RESULTS AND DISCUSSION

3. 1 Photolytic decomposition of ozone and catalytic activity of TiO₂ (P25)

Figure 2 illustrates the time course for ozone decomposition. We tested O_3 decomposition under illumination without photocatalyst. The O_3 concentration and flow rate at the inlet of the reactor containing a bare glass plate (without TiO_2) were 1.0 ppm and 1.0 l/min respectively and ultraviolet (UV) wavelength was 300 \sim 400 nm.

When only illumination was given, O_3 concentration was decreased by 3.1% with black light (300~400 nm). Ohtani *et al.* (1992) have reported that even in the absence of TiO_2 , irradiation of light with wavelengths >200 nm induced the O_3 decomposition of 8%. However, irradiation of light with wavelengths >290 nm did not induce O_3 decomposition. Therefore, decomposition corresponds to the photolysis of gaseous O_3 in the wavelength region of 200~290 nm (Griggs, 1968). As the spectrum of the black light lamp used in this experiment showed weak emission in UV region (<290 nm), it seems that photolytic decomposition of gaseous ozone occurs under our experimental condition.

$$O_3 + hv \rightarrow O_2 + O \cdot \tag{1}$$

$$O \cdot + O \cdot \rightarrow O_2 \tag{2}$$

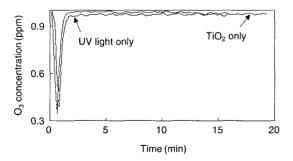


Fig. 2. Representative time course of ozone decomposition in the presence of TiO₂ only or UV light only (initial ozone conc.: 1.0 ppm, flow rate: 1.0SLPM, lamp: black lamp (300~400 nm), relative humidity: 0%, TiO₂ coated on glass plate: 1.0 mg/cm²).

A heterogeneous catalytic reaction consists of the following processes: 1) mass transfer of reactants to the catalyst surface, 2) adsorption of reactants on the surface, 3) chemical reaction on the surface, 4) desorption of products from the surface, and 5) mass transfer of products from the surface into the bulk of the fluid.

3. 2 Influence of loaded TiO₂ (P25) weight on photocatalytic decomposition of O₃

Figure 3 shows the relation between O₃ removal ratio and loaded weight of catalyst for a selected reaction condition at the same irradiation condition when P25 was used as photocatalyst. Under fixed light intensity $(0.49 \text{ mW/cm}^2 \text{ at } 300 \sim 400 \text{ nm})$, almost the same removal ratio was obtained regardless of the amount of TiO₂ in the range from 0.5 mg/cm² to 4 mg/cm², indicating that a feature of the reaction rate above 0.5 mg/ cm² is constant regardless of the loaded weight of photocatalyst. This is thought to be due to the fact that the number of photon adsorbed by photocatalyst is limited because light could penetrate into only certain depth of photocatalyst. Ohtani et al. (1992) suggested that the photocatalytic activities of TiO2 powder are independent of their surface area. As shown by data below 0.5 mg/cm² in Figure 3, a loaded weight of photocatalyst becomes smaller, so does ozone removal ratio. Therefore, in this study, the amount of photocatalyst loaded on glass plate was fixed to 1.0 mg/cm² for each experiment.

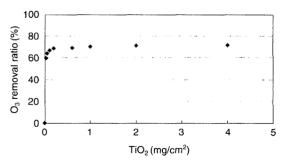


Fig. 3. The relation between ozone removal ratio and loaded amount of photocatalyst (initial ozone conc.: 1.0 ppm, flow rate: 1.0SLPM, lamp: black light (300~400 nm), relative humidity: 0%).

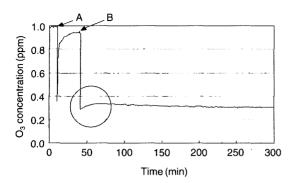


Fig. 4. Representative time course of photocatalytic decomposition of ozone in the presence of TiO₂ (initial ozone conc.: 1.0 ppm, flow rate: 1.0SLPM, lamp: black lamp (300~400 nm), relative humidity: 0%, TiO₂ coated on glass plate: 1.0 mg/cm², A: reaction start, B: UV irradiation start).

Figure 4 illustrates the typical time course for photocatalytic decomposition of O_3 on TiO_2 in this study. At first, in the presence of TiO_2 catalyst, a decrease in O_3 concentration was observed in the dark. However, this may not be due to the decomposition in the dark but due to the adsorption by TiO_2 . When the outlet O_3 concentration becomes about 90% of initial one, UV light was irradiated (titled B in Fig. 4). As shown in Figure 4, O_3 was removed by UV irradiation quickly. We found that outlet O_3 concentration gradually rised after a rapid decrease and then stabilized as shown in circle in Figure 4. A desorption rate of O_3 adsorbed in TiO_2 in the dark may be higher than the rate of O_3 photocatalytic decomposition at the beginning of irradiation.

3. 3 Influence of initial concentration on photocatalytic decomposition of O₃

The influence of O_3 concentration on the kinetic oxidation rate of O_3 over TiO_2 is shown in Figure 5. As the initial O_3 concentration becomes higher, the O_3 oxidation rate becomes higher.

In general, for gas-solid reaction, the kinetics would follow the Langmuir-Hinshelwood (L-H) model, in which the reaction rate varies proportionally with the surface coverage (θ) as:

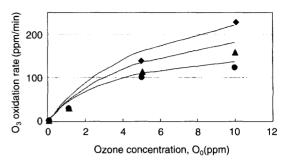


Fig. 5. Influence of initial ozone concentration on the ozone oxidation rate over TiO₂ (flow rate: 1.0SLPM, lamp: black light (300 ~ 400 nm), relative humidity: 0%, the amount of TiO₂ coated on glass plate: 1.0 mg/cm², photocatalysts: P25 (diamond), ST-01 (circle), E-23 (triangle up).

$$r = k \Theta = \frac{kKC}{(1 + KC)} \tag{3}$$

Where k is the rate constant, related to the limiting rate of reaction at maximum coverage for the experimental conditions; K is the adsorption equilibrium constant and reflects the proportion of solute molecules which adhere to the catalyst surface; and C is the concentration of the O_3 . Since the rate expression (Eq. 3) is nonlinear, least-squares optimization was used to determine values for the constants. The results of the optimization are given in Table 1. The rates from Eq. (3) are in good agreement with the measured data as seen in Figure 5.

After substituting this rate expression into a mass balance on a plug flow reactor, the following expression is obtained (Zhang *et al.*, 1994):

$$\frac{V}{Q} = \frac{1}{kK} \ln\left(\frac{C_0}{C}\right) + \frac{1}{k} (C_0 - C) \tag{4}$$

Where V is the volume of the reactor, Q is the flow rate through the photoreactor and C_0 is the initial concentration of O_3 . The quantity, V/Q is known as the contact time (residence time) for the reaction or average time that a molecule passes through the reactor.

The linear characteristic of Eq. (4) was used for fitting the experimental data after changing the form as

Table 1. Langmuir-Hinshelwood parameters obtained for the photocatalytic decomposition of gaseous

Catalysts	k (ppm/min)	<i>K</i> (1/ppm)	<i>kK</i> (1/min)
P25	357.14	0.16	58.14
ST-01	277.78	0.19	52.63
E-23	185.19	0.29	53.76

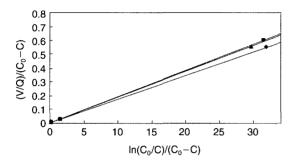


Fig. 6. Langmuir-Hinshelwood plots for ozone decomposition (TiO₂: P25 (diamond), ST-01 (square), E-23 (triangle up).

follows:

$$\frac{V/Q}{(C_0 - C)} = \frac{1}{k} + \frac{1}{kK} \frac{\ln(C_0 - C)}{(C_0 - C)}$$
 (5)

If one assumes that L-H kinetics hold for a plug flow reactor, then a plot of $(V/Q)(C_0-C)^{-1}$ vs. $\ln (C_0/Q)$ $C(C_0 - C)^{-1}$ should be linear. Eq. (5) can be tested using different values of C_0 and C. Here the initial concentration of O₃ was varied. Figure 6 shows the experimental data obtained for O3, which are in good agreements with this integral rate-law analysis. The kinetic parameters, k and K were obtained using linear least squares analysis. Table 1 summarizes the kinetic parameters of photocatalytic decomposition for O₃ under the identical conditions. It is interesting to note that the photocatalytic degradation rate is related to k and K; therefore, a higher adsorption constant does not always result in a higher reaction rate. Although it could be thought that a high rate constant leads to a rapid reaction, the reaction rate based on the L-H kinetic model

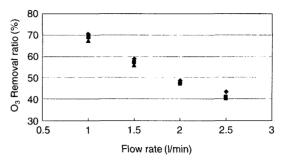


Fig. 7. The influence of flow rate on photocatalytic decomposition of gaseous ozone (initial ozone concentration: 1.0 ppm, lamp: black light (300 ~ 400 nm), relative humidity: 0%, the amount of TiO₂ coated on glass plate: 1.0 mg/cm², TiO₂: P25 (diamond), ST-01 (square), E-23 (triangle up).

depends simultaneously on k and K. In our experiment, the rate constant did increase in the order of P25, ST-01 and E-23, but the multiplied products of k and K did increase in the order of P25, E-23 and ST-01.

3. 4 Influence of flow rate on photocatalytic decomposition of O₃

Figure 7 shows that the removal ratio of O_3 decreases linearly with the increasing flow rate for each catalyst. This is expected because a better gas-solid contact can be obtained with lower flow rate.

The photocatalytic reactor used in this study has negligible intraparticle transport resistance due to its non-porous catalyst configuration. Thus, only the bulk fluid-to-catalyst diffusion mass transfer is available. The effect of external mass transfer limitations which may occur in transfer of the gas from bulk to the exterior surface of the TiO₂ was investigated using different flow rates of the gas.

Figure 8 shows the dependency of O_3 reaction rate on the gas flow rate for each catalyst under the relative humidity of 0%. For the system used in this study, gas velocity at each flow rate is as follows; 0.067 m/s (1 l/min), 0.1 m/s (1.5 l/min), 0.133 m/s (2 l/min) and 0.167 m/s (2.5 l/min). The results in Figure 8 show that the reaction rate increases with the flow rate.

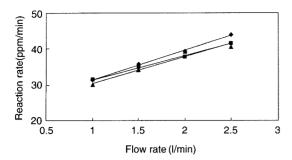


Fig. 8. The influence of flow rate on photocatalytic oxidation rate of ozone over TiO₂ (initial concentration: 1.0 ppm, lamp: black light (300~400 nm), relative humidity: 0%, TiO₂: P25 (diamond), ST-01 (square), E-23 (triangle up).

3. 5 Influence of relative humidity

In aqueous solution, water molecule takes part in the radical chain reaction as follows:

$$TiO_2 + hv \rightarrow h^+ + e^- \tag{6}$$

$$O_3 + e^- \rightarrow O_3^- \tag{7}$$

$$O_3 + H^+ \to HO_3 \tag{8}$$

$$HO_3 \rightarrow O_2 + OH \cdot$$
 (9)

$$OH \cdot +O_3 \rightarrow O_2 + HO_2 \tag{10}$$

That is, water molecule enhances the consumption of O₃. But, for gaseous O₃ decomposition, the experimental result (Figure 9) shows that water molecule decreases the O₃ decomposition percentage for each photocatalyst because higher humidity evokes low decomposition percentage. It has been reported that during O₃ decomposition over TiO₂, O₃⁻ and O₄⁻ were detected as intermediates (Gonzalez–Elipe *et al.*, 1981). Hoffmann *et al.* (1995) and Ohtani *et al.* (1992) reported that the following reaction sequence may occur on the surface of TiO₂:

$$TiO_2 + hv \rightarrow h^+ + e^- \tag{11}$$

$$\equiv \text{TiOH} + e^- \rightarrow > \text{TiOH}^- \tag{12}$$

$$O_3 + > TiOH^- \rightarrow O_3^- + \equiv TiOH$$
 (13)

$$O_3^- + O_3 \rightarrow O_4^- + O_2$$
 (14)

$$\equiv \text{TiOH} + \mathbf{h}^+ \to \equiv \text{TiOH}^+ \tag{15}$$

$$\equiv \text{TiOH}^+ + \text{O}_4^- \rightarrow \text{TiOH} + 2\text{O}_2 \tag{16}$$

$$H_2O \rightarrow OH^- + H^+ \tag{17}$$

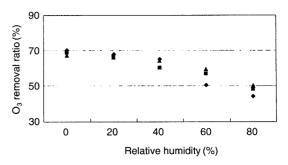


Fig. 9. The influence of relative humidity to photocatalytic decomposition of gaseous ozone (flow rate: 1.0SLPM, initial ozone concentration: 1.0 ppm, lamp: black light (300~400 nm), TiO₂: P25 (diamond), ST-01 (square), E-23 (triangle up).

$$O_2 + e^- \rightarrow O_2^- \tag{18}$$

$$O_{2}^{-} + H^{+} \rightarrow OOH \tag{19}$$

$$OH^- + h^+ \to OH \tag{20}$$

Equations (11) \sim (16) depict the decomposition reactions for O_3 , and Equations (17) \sim (21) depict the processes in which water molecule scavenges e^- (exited electron) and h^+ (positive hole). It is obvious that water molecule competes with O_3 for scavenging exited electron and positive hole on the surface of TiO_2 . So the humidity decreases the O_3 removal ratio. This is proved by our experimental results shown in Figure 9.

4. CONCLUSIONS

In this study, the characteristics of heterogeneous photocatalytic decomposition of low concentration level of O₃ over TiO₂ were investigated for various operating parameters. The conclusions from the present work can be summarized as follows:

- 1. Under fixed light intensity, flow rate, initial O_3 concentration, and TiO_2 loaded amount in the range from 0.5 mg/cm² to 4 mg/cm² gained almost the same O_3 removal ratio and the ratio became constant for more amount of catalysts in this reactor.
- 2. For the influence of initial O_3 concentration to the photocatalytic decomposition of gaseous O_3 , higher

- initial O_3 concentration resulted in greater oxidation rate of O_3 , and the experimental data was kinetically in good agreement with Langmuir-Hinshelwood kinetic model.
- 3. The decomposition of O₃ decreased with the increasing flow rate. This is expected because a better gas –solid contact can be obtained with lower flow rate.
- 4. It was observed that water molecule competes with O_3 for scavenging exited electron and positive hole on the surface of TiO_2 catalyst. Therefore, the O_3 removal ratio was decreased with increasing relative humidity.

According to our results, photocatalytic decomposition of gaseous O₃ over TiO₂ thin film has the high potential to be used for purification of vent gases from industrial process and indoor/outdoor air.

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